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Interacting fields approach for evolving spatial phenomena: application to erosion simulation for optimized land use

We propose a computational framework and strategies for performing tasks necessary for evaluation and optimization of land use management within an advanced GIS modeling environment. Such tasks involve modeling of landscape processes, simulation of impact of human activities on these processes and optimization of preventive measures aimed at creating sustainable landscapes. A typical example of interaction between landscape processes and human activity is water erosion, a natural process which can be accelerated or greatly reduced by human intervention. In our implementation, the underlying 2D water and sediment transport equations are solved within approximate diffusive wave hydrologic and detachment/transport capacity models using stochastic methods (Monte Carlo). The methods employ distributed input parameters and enable us to simulate the impact of complex terrain, various soils and land cover changes on the spatial distribution of erosion and deposition. The presented approach together with land use optimization scenario is illustrated by an application to an experimental farm in Germany, carried out within 3D dynamic GIS environment.

This paper, including color images and animations, can be found at the following URLs:

http://www.ncgia.ucsb.edu/conf/sf_papers/latest_copies6/mitas_lubos2.0/mitas.html

<http://www.cecer.army.mil/grass/viz/SF.final/mitas.html>

This document requires a browser which supports tables.

The full size figures and animations can be retrieved by clicking on reference images.

INTRODUCTION

Efforts to balance the economic development with environmental protection increase the demand for simulation tools enabling predictions of the human impact on landscape. In order to prevent irreversible changes and avoid costly, ineffective solutions the simulation tools should provide detailed spatial and temporal distributions of modeled phenomena. Statistical averages for the entire study areas or predictions only for a certain point, such as watershed outlet are often insufficient. New developments in GIS, especially support for multivariate temporal data processing, analysis, and visualization (Mitasova et al. 1995, GRASS4.2 libraries) make such simulations possible and GIS plays an important role in the development and applications of distributed landscape process models (e.g., Engel 1995(review with links), Vieux 1995, Saghafian 1995, Leavesley et al. 1995).

The goal of our paper is to outline a methodological framework for distributed models based on the solution of the "first principles" master equations for multi-variate fields and use these tools for the

distributed landuse scenarios optimization. The basic platforms of our approach are described as follows:

- **Representation of phenomena as multivariate fields (i.e. as genuine distributed objects).** This advances lumped or semi-lumped description into the high resolution distributed one with advantages for spatial and temporal analysis. Such approach requires multivariate tools which deal efficiently with transformation from one discrete format to another, and with the possibilities of continuous representation for calculation of gradients, curvatures and other quantities. For these purposes we use the multivariate splines developed previously (Mitasova et al. 1995) and tested on a variety of 2D, 3D and 4D data (Table 1.).
- **Phenomena description and prediction based on solving master equations which determine the configuration or evolution of corresponding fields.** This enables us to perform simulations which are formulated in terms of fundamental physical processes such as flux, diffusion, etc., and we can use the known mathematical and physical apparatus to analyze the solutions. This level of rigour avoids vague concepts which often characterize various semi-empirical schemes and allows us to clearly specify inputs, governing parameters and to understand the character of processes involved as well as final outputs. We can also build upon experience from other disciplines in using efficient and robust methods for solving the underlying master equations.
- **Formulation of cost functionals and their optimizations.** The proper quantification of objectives is very important for an efficient solving of desired landuse practices. It is now well understood that the formulation of an appropriate cost or objective functional provides a powerful strategy how to deal with this task. The cost functional depends on the fields and includes also various conditions or restrictions. The human actions can change the character or properties of some of the input fields or modify their future evolution. The "space of human actions" is then explored for estimating the optimal solutions.

We illustrate these three fundamental concepts on a case of erosion prevention by optimized landuse scenario. Water erosion, with its economical and ecological impact, is a typical example of problems targeted by our approach. It is a genuine space-time distributed phenomenon with several natural components such as terrain, soils, cover and climate effects. These can be naturally and elegantly represented by multivariate fields (terrain surface, terrain cover, water and sediment distributions, water and sediment fluxes, soil distribution, etc). The erosion processes can be described by master continuity and momentum conservation equations. Some of these fields such as land cover can be influenced or changed by human intervention and therefore will naturally enter the optimization process.


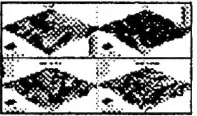





In the following sections we describe the **working examples of advanced methods and approaches** which enabled us to start from basic input data (terrain, cover, soils, etc) and get to the actual creative process of landuse optimization in a fully distributed manner.

LANDSCAPE CHARACTERIZATION IN 3D SPACE AND TIME

The basic objects which enter into distributed models of landscape processes are given by functions which depend on the position in 3D space and time: *multivariate scalar and vector fields*. These fields represent various phenomena such as terrain, soil properties, land cover, fluxes of matter. They are usually represented in a GIS database in a discrete form as sets of points (sites, lines or polygons) or rasters. However, they can be transformed to continuous representations using expansions in an

appropriate basis set such as multivariate regularized spline with tension (Mitasova et al. 1995), as illustrated by examples in Table 1. or by Hargrove et al. (1995). Such representation is not restricted to continuous fields, an example of effective handling of surfaces with faults using splines and GIS tools was developed by Cox et al. 1994. Phenomena represented by classes, such as vegetation or land use, can be represented as fields with faults as well (Table 1., land cover).

Table 1. Examples of landscape phenomena representations by multivariate fields
(Click on the image to retrieve full size picture or animation)

Representations of natural phenomena as multivariate fields modeled by bi-, tri-, and quad-variate splines		
phenomenon (field)	3D (dynamic) "map"	GIS format (discrete representation)
elevation		points (x,y,z), lines (contours), 2D raster (DEM)
elevation gradient and curvatures		2D rasters derived from $z=f(x,y)$
precipitation		points(x,y,z,p,t), 2D raster time series
soil horizons		points (x,y,z,w), 2D raster vertical series
land cover		polygon, raster
underground concentrations of chemicals		points (x,y,z,t,w), 3D raster time series
concentration of chemicals in water		points (x,y,z,t,w), 3D raster time series

The fields are static or evolving in time. The evolution can be monitored and data from monitoring can be stored in a GIS database in various formats, for example, as time series of sites which can be further transformed to time series of 2D or 3D rasters (e.g., Table 1.,chem. concentrations). To predict the future states of these fields, we need to understand the processes controlling their evolution and formulate models simulating their fundamental behavior. The GIS software is being enhanced to support

such simulations by providing the adequate data structures, including 2D and 3D floating point raster data (Waupotitsch and Shapiro 1995), multidimensional site data (McCauley 1995), support for temporal data (Brown and Shapiro 1995), methods for transformation between discrete and continuous representations using multivariate interpolation by radial basis functions (Mitasova et al. 1995), and tools for interactive multidimensional dynamic cartography (Brown et al. 1995).

FLUXES IN LANDSCAPE

The problem of landscape process simulations related to land use change have been attacked by a variety of approaches such as rule-based models, cellular automata, probability transitions cell based-models (Berry et al. 1995) or as continuous fields and differential equations (Maidment 1995). In our approach to modeling of erosion processes we start from the continuous formulation and we use the fundamental conservation laws (matter and approximate momentum conservation).

Water flux

One of the primary processes in landscape, influencing the distribution of soils, plants and people is water flow. While there are numerous empirical and process based models for modeling the waterflux in streams and rivers, spatial distribution of water on complex hillslopes (crucial for soils and organisms) is still being modeled by rather rough estimates unable to provide sufficient detail for modeling of some important water related processes, such as erosion.

In general, the overland flow is described by the continuity equation:

$$\frac{\partial h(\mathbf{r}, t)}{\partial t} = i(\mathbf{r}, t) - \nabla \cdot \mathbf{q}(\mathbf{r}, t) \quad (1)$$

$$\mathbf{q}(\mathbf{r}, t) = h(\mathbf{r}, t)\mathbf{v}(\mathbf{r}, t) \quad (2)$$

where h is the water depth [m], i is the rainfall excess=rainfall-infiltration [m/s], \mathbf{q} is the water flux [m.m/s], \mathbf{v} is the flow velocity [m/s], $\mathbf{r}=(x,y)$ is the position [m], and t is the time. The velocity \mathbf{v} is related to h by Mannings or Chezy law and the momentum conservation equation (Haan 1994).

Current GISs provide the tools for the simplest approximate solution of equation (1) for steady state flow and constant velocity, based on the upslope contributing area. There are numerous algorithms available for its estimation, such as D8 (Figure 1a), or an improved approach based on the vector-grid algorithm (Mitasova et al. 1995, Figure 1b). While these geometric approaches provide enough information for a wide range of applications (Moore et al. 1992), they become problematic when applied to areas with spatially variable cover and complex terrains with significant spatial variations in flow velocity.

A more realistic approximation which takes into account the flow velocity is based on the steady state solution of (1) in the 2D kinematic wave (Figure 1c) and 2D approximate diffusive wave (Figure 1d) (Mitas and Mitasova, in prep.). The 2D approximate diffusive wave solution can be found by solving the steady state of (1) which can be written in the operator form as:

$$W[h^{5/3}(\mathbf{r})] = -\frac{e}{2} \nabla^2 [h^{5/3}(\mathbf{r})] + \nabla \cdot [h(\mathbf{r})\mathbf{v}(\mathbf{r})] = i(\mathbf{r}) \quad (3)$$

where e is the diffusion constant and W is the operator. We use a stochastic method to solve the equation (3). This approach is based on the representation of the solution h by a large set of random walkers

(sampling points) which are propagated according to the Green's function corresponding to the inverse operator W . The diffusion term in (3) describes (approximately) a backwater effect and also helps to reduce the artificial features in water surfaces on hillslopes (Figure 1a,b,c) caused by flowtracing on a regular discrete grid, which make the application of such water surfaces in erosion/deposition models problematic. This approach can be extended also for non-stationary event based modeling.

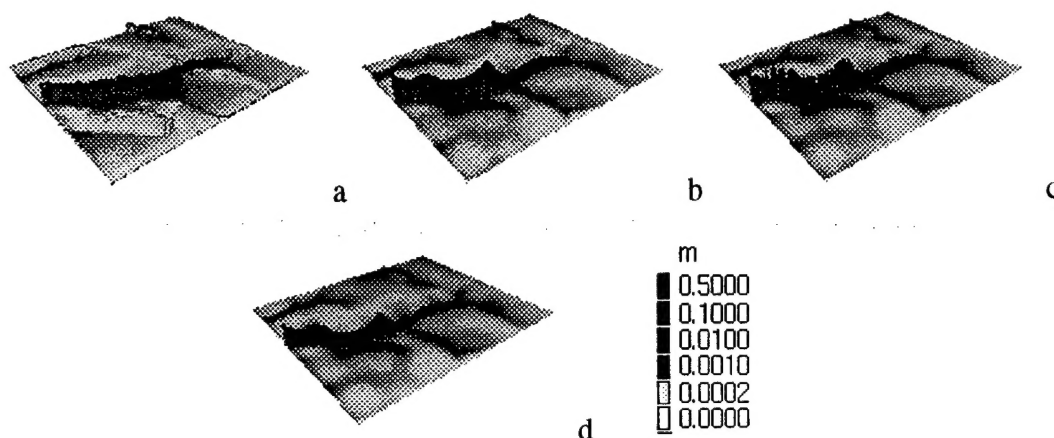


Figure 1. Steady state water depth estimated by a) upslope contributing area using D8 algorithm, b) upslope contributing area using vector-grid algorithm c) 2D kinematic wave approximation, d) 2D approximate diffusive wave

Other approaches which solve for temporal and spatial distribution of water depth during storm events are, for example, kinematic (Garrote and Brass 1995, Vieux 1995) or diffusive wave models (Saghafian 1995, Figure 1e), already integrated with GIS (r.water.fea, r.hydro.CASC2d). The choice of hydrology model complexity and realism depends on the type of application and for our current efforts steady state provides adequate information for assessing the impact of water flow on landscape at the time scale of days to years. The steady state solutions are also consistent with the new generation erosion model Water Erosion Prediction Project (WEPP) (Flanagan and Nearing 1995).

Sediment flux and erosion/deposition in complex terrain

Soil erosion involves detachment, transport and deposition. The interaction between soil detachment and sediment transport is controlled by water flux, terrain, soil and cover. This interaction is very difficult to capture by traditional empirical models or models based on the geometrical analysis of terrain. While some of these models provide adequate tools for a qualitative assessment of erosion risk for large areas with complex terrain (Moore and Wilson 1992, Mitsova et al. 1996), they are insufficient for modeling of impact of spatially variable landuse and simulation of erosion protection measures effectiveness.

The basic relationship for fundamental erosion processes is continuity of mass. For erosion by 2D overland flow, the continuity equation is (Foster and Meyer 1972, Govindaraju 1991)

$$\frac{\partial [r_s(r, t)h(r, t)]}{\partial t} + \nabla \cdot \mathbf{q}_s(r, t) = \text{sources} - \text{sinks} \quad (4)$$

where $\mathbf{q}_s = r_s c(r, t) \mathbf{q}$ is the sediment flux [kg/(ms)], c is the sediment concentration [particle/(m.m.m)], and r_s is the mass per sediment particle [kg/particle]. Further:

$$\text{sources} - \text{sinks} = D_r(\mathbf{r}, t) + D_i(\mathbf{r}, t) = C[T(\mathbf{r}, t) - |q_s(\mathbf{r}, t)|] + D_i(\mathbf{r}, t) \quad (5)$$

$$T(\mathbf{r}, t) = K_t[r_w g h(\mathbf{r}, t) S(\mathbf{r}) - \tau_{cr}] \quad (6)$$

where T is the sediment transport capacity [kg/(ms)], D_r is the rill erosion or deposition rate, D_i is the interrill contribution [kg/(m.m.s)], C is the first-order reaction coefficient dependent on soil and cover [1/m], r_w is the mass density of water [kg/m.m.m], $S(\mathbf{r}) = \text{grad } z(\mathbf{r})$ is the slope [m/m], $z(\mathbf{r})$ is the elevation [m], K_t is the transport capacity coefficient, $g = 9.81$ is gravitational acceleration [m/s.s.s]. We assume that the critical shear stress τ_{cr} is negligibly small.

Rill detachment and deposition are proportional to the difference between transport capacity and sediment load (eq. 5). This relationship defining the interaction between sediment load and transport capacity (Foster and Meyer, 1972) is based on a stream power concept (Haan 1994) and can be expressed as:

$$D_r/D_c + q_s/T = 1 \quad (7)$$

where $D_c = CT = K_d r_w g h(\mathbf{r}) S(\mathbf{r})$ is the detachment capacity and K_d is the detachment capacity coefficient (rill erodibility).

We solve the continuity equation in 2D form for steady state water flux with small diffusive term with amplitude d , rewritten in the operator form as:

$$\mathcal{L}q(\mathbf{r}) = -\frac{d}{2} \nabla^2 q(\mathbf{r}) + \nabla \cdot [q(\mathbf{r}) \mathbf{v}(\mathbf{r})] + Cq(\mathbf{r}) |\mathbf{v}(\mathbf{r})| \quad (8)$$

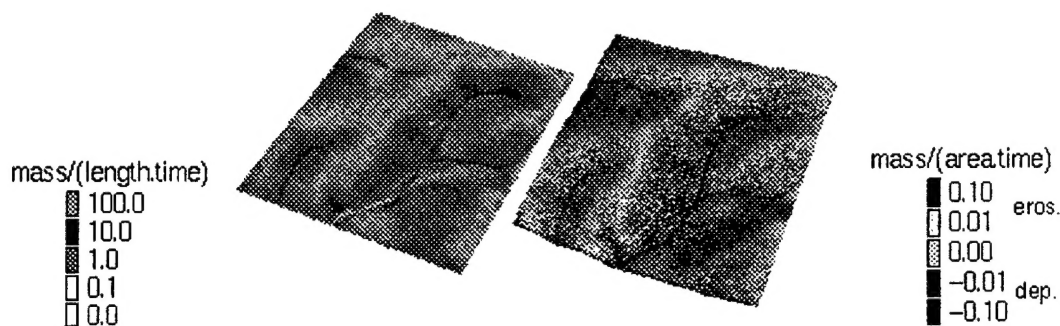
and then the erosion equation is

$$\mathcal{L}q(\mathbf{r}) = CT(\mathbf{r}) + D_i(\mathbf{r}) \quad (9)$$

where $\rho = r_w c(\mathbf{r}) h(\mathbf{r})$. The interpretation of the equation (8) is clear: the first term is the diffusion (which in our case is very small, represents the smoothing component of the soil transport), the second term is the drift driven by the water velocity $\mathbf{v}(\mathbf{r})$ and the third term is the 'potential' which is dependent on the velocity magnitude: the larger the velocity, the smaller the concentration of sediment.

Net erosion and deposition is then estimated as a divergence of sediment flux. Further details about this approach and comparisons with previous estimations of erosion/deposition by the directional derivative (Mitasova et al. 1995, Wilson and Moore 1992) will be given elsewhere Mitas and Mitasova, (in prep.).

The equation (8) is solved analogically as the equation for water flux using a stochastic method (Mitas and Mitasova, in prep.), illustrated by Movie 1.:



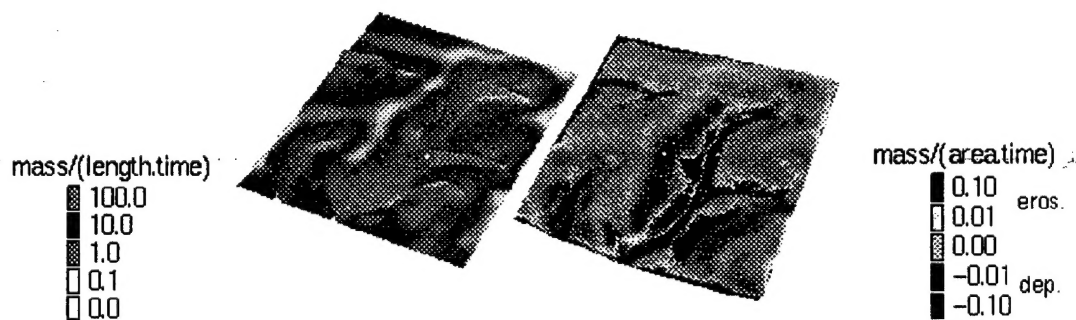
Movie 1. Solution of erosion equations by Monte Carlo, illustrated by a surface representing the sediment flux and by terrain with draped erosion/deposition

The animation shows that the approximate estimation of sediment flux is reached for a relatively small number of Monte Carlo sampling points. Accurate calculation of erosion/deposition estimated by derivatives of sediment flux, which are very sensitive to statistical noise, can be estimated to a given accuracy either by employing large number of walkers or by smoothing out the noise numerically.

Influence of uniform soil and cover parameters

In the erosion model the influence of soil and cover is represented by the following basic parameters: Mannings n , detachment rate coefficient (erodibility) Kd , and sediment transport coefficient Kt . These coefficients are functions of soil and cover properties such as soil texture, canopy, roots, management practices etc., and their estimation and development of various adjustment factors is described in WEPP (Flanagan and Nearing 1995). Constants for estimation of detachment and sediment transport capacity are still under development and detailed discussion of these parameters is beyond the scope of this paper. However, we will use the following examples to elucidate the role of these parameters in modeling spatial distribution of areas with erosion or deposition.

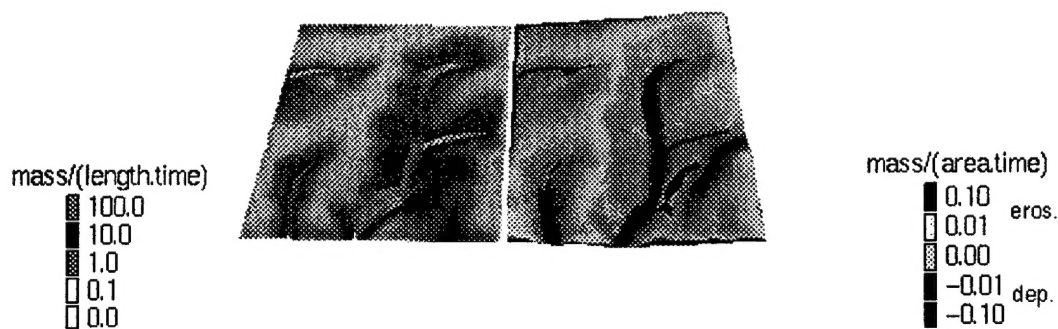
The most important parameter controlling the *borderline between the erosion and deposition* areas is the first-order reaction coefficient C related to the ratio of detachment capacity and sediment transport capacity. In the first example, we simulate the situation when the study area has constant transport capacity coefficient Kt but detachment capacity coefficient Kd increases, so that the ratio C increases from 0.0005 to 100 (Movie 2).



Movie 2. Change in the spatial distribution of sediment flux and erosion/deposition due to the change in C with increasing Kd and $Kt=const.$

For small values of C (e.g., clay with very low fall velocities and low detachment capacity) water has the power to carry almost all detached sediment into the stream. For values $C > 1$ (e.g., sandy soils with relatively high fall velocities which detach easily), the sediment flux quickly reaches the sediment transport capacity and deposition occurs relatively high in the hillslope. This is the case of transport limited erosion/deposition modeled by a simplified approach described e.g. by Mitsova et al. 1995 and Moore and Wilson 1992. For both cases the magnitude of sediment flux in the stream remains the same while the distribution of erosion and deposition over the landscape changes significantly. This simulation is a good example which shows that calibrating the erosion model using only the observed values of sediment flux at an outlet does not guarantee correct predictions of erosion/deposition on the complex hillslopes within the watershed.

Change in Kt while Kd is constant also changes the spatial distribution of erosion/deposition, however there is a big difference in the amount of sediment delivered to streams. For $Kt \ll Kd$ and $C \gg 1$ most of the detached material deposits before it enters the stream, for $Kt \gg Kd$ and $C \ll 1$, there is only a very small deposition and most of the detached material is delivered into streams (Movie 3). This example illustrates how the potential changes in soil properties and cover which increase transport capacity can trigger severe erosion.



Movie 3. Change in the spatial distribution of sediment flux and erosion/deposition due to the change in C with increasing Kt and $Kd = \text{const.}$

If $C = \text{const.}$ and Kt , Kd change with the same rate, the spatial distribution of erosion/deposition is the same, and only its magnitude changes (Figure 2.)



Figure 2. Change in the magnitude of erosion/deposition with $C = 1$ and increase in both Kt and Kd ;

a) $n=0.1$, $K_d=K_t=0.0003$ (grass on sandy soil); b) $n=0.05$, $K_d=K_t=0.03$ (bare sandy soil). Surface topography represents the sediment flux, color is the erosion/deposition.

It is important to note that the parameters K_t and K_d which are dependent on the soil and cover properties are interrelated and the change in one parameter is usually accompanied with the change in the second parameter. As we have demonstrated, it is their ratio, which plays important role in the spatial distribution of erosion and deposition. With better understanding of the physical basis of these parameters the analysis outlined above can be used for identification of those soil and cover properties which can be targeted for the most effective erosion prevention.

Erosion/deposition for spatio-temporal changes in soil and cover parameters

Erosion process is highly dynamic and its temporal variability can be modeled at various time scales from minutes (event based models such as ANSWERS or AGNPS), days (WEPP), to geological time (Moglen and Brass 1994). For land use management applications we adapt the concept used in WEPP and we simulate erosion under steady state flow for variable climatic, soil and cover parameters as they change during the year.

We have used data from the experimental farm of the Technical University in Munchen in Germany (courtesy Dr. Karl Auerswald) such as elevation, soil core data and current land use and we simulated erosion/deposition under various land cover and climatic conditions, using simulated data from WEPP.

Comparison of results with the spatial distribution of colluvial deposits (Figure 3) and with the pattern of linear erosion after a 150 year storm (Figures 3, 4) indicates, that for this area with mostly sandy soils, the terrain controls the long term spatial pattern of deposition reflected in colluvial deposits observed in mostly concave areas (Figure 3a, Movie 4). Land cover has more significant impact on the magnitude of erosion and short term linear erosion features (Figure 4).

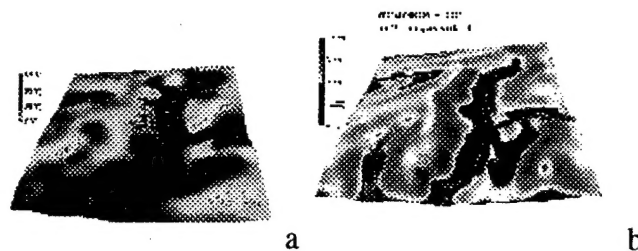


Figure 3. Comparison of a) observed spatial distribution of colluvial deposits (depth in cm) and linear erosion features after the 150 year storm (red lines) with b) simulated erosion/deposition with homogeneous bare soil conditions.



Figure 4. Simulated a) sediment flux and b) erosion/deposition with the incorporation of the a) current

land use influence.



Movie 4. Slicing through colluvial deposits

We have also simulated the changes of erosion/deposition during the year due to the changes in rainfall and land cover illustrated by Movie 5. and Figure 5.



Movie 5. Change in the land cover due to the plant growth and harvest

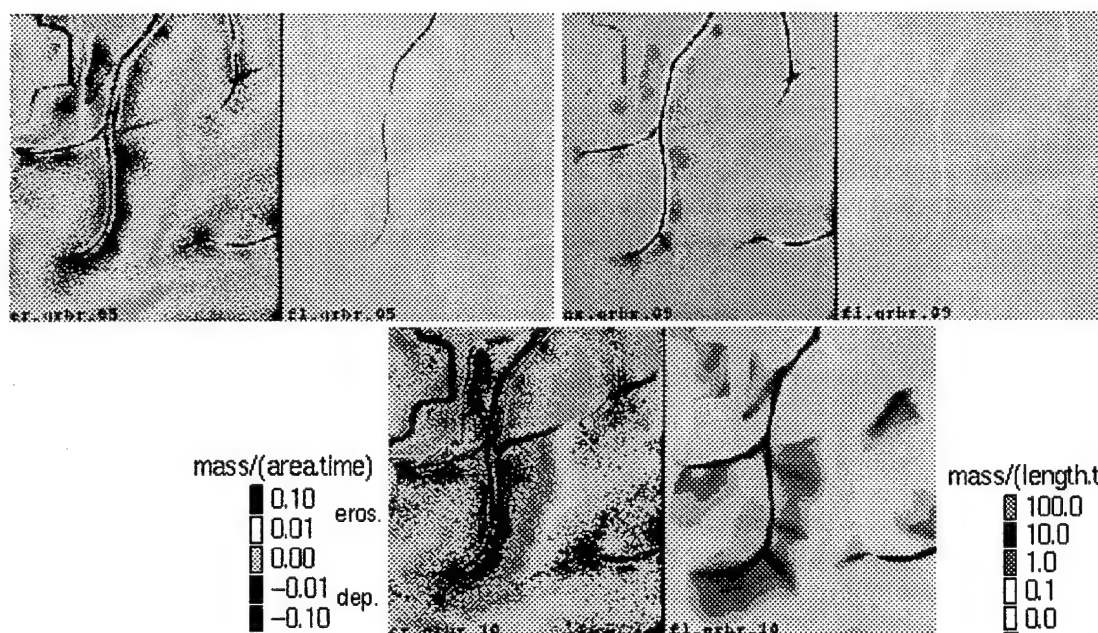


Figure 5. Impact of changes in rainfall and cover on erosion/deposition distribution and sediment flux in May, September and October.

The highest risk of erosion was predicted in October, when there is minimum cover and enough rainfall to produce significant runoff. Although May had the most intense storm, the erosion was lower due to the good cover provided by both the grass area and the agricultural area with winter wheat.

LANDUSE OPTIMIZATION

Human activity changes character or properties of landscape components (e.g. cover) which can be represented as an appropriate change in the corresponding field. These changes influence the natural phenomena through various interactions which can be described by the governing master equations.

Well-defined and quantified impact of human actions on nature enable researchers to formulate objectives or costs in order to either predict the future development or, more importantly, to achieve a desired sustainable development. The desired objectives can include, e.g., maximization of land use for production or military training with minimized impact on environment, or prevention of unacceptable changes in environment in the given time horizon with minimized costs (Johnston and Hopkins 1994). Because of the extremely complex nature of the problem, the optimization tasks are often out of the scope of ordinary techniques as they involve multivariate fields (possibly evolving in time) and also because of the special type of human action (like instantaneous point sources such as contamination which spreads out in the time horizon of a few years or clear cut of a forest with consequent erosion). Therefore this problem requires a formulation of general methods which can deal with complicated types of "configuration or state spaces". For our case we can define the *state space* as a set of fields (i.e., a particular set of multivariate functions) which describe components of the studied phenomenon. Available information and models such as initial fields values, are provided by a GIS and are used as inputs into the master equations and their solvers. In addition, we need to express the objective (cost) functional which is to be minimized within given constraints. The constraints can be formulated in the form of "external" fixed influences (e.g. part of land cover which cannot be changed), thresholds on evolving fields (e.g. erosion beyond certain level is unacceptable) and so forth. The general form of the cost functional can be given as:

$$I = \int \int dr dt F(\{z_i(r, t)\}) \quad (10)$$

where I is the cost functional, z_i are the input spatio-temporal fields, F is a function which determines of cost for a particular set of z_i (point in the state space). In general, a minimization of (10) can be a very complex task. In order to carry out the minimization of the functional (10) we have to define the following:

- "distance" in the state space
- efficient representation of the fields which can be varied by the human impact (e.g. using appropriate basis function expansions)
- "movement" of the space of field configurations

Another important task is to formulate efficient minimization strategies. Because we are dealing with a multivariate problem which often involves non-linearities the cost functional can have many local minimas. This requires use of robust minimization methods such as simulated annealing or genetic algorithms.

A simple example of using GIS and the erosion model to optimize land use by finding a more effective spatial distribution of protective grass cover while keeping the ratio between the agricultural area and area protected by grass constant is illustrated in Figures 6 and 7. Under the current land use (Figure 6a), there is still a significant amount of sediment delivered to the stream (Figure 6b) with strong potential of creating rills and gullies (Figure 6c, dark red) which is in agreement with observations of big storm effects, presented in Figure 4. Redesigning the land use so that the protective grass cover is located in the highest erosion risk areas (Figure 7a) can dramatically reduce soil loss and sediment delivery to the streams (Figure 7b,c). The crest in sediment flux in areas with observed gullies disappears and is replaced by light deposition caused by the decrease in water velocity in the grass strip (Figure 7c).

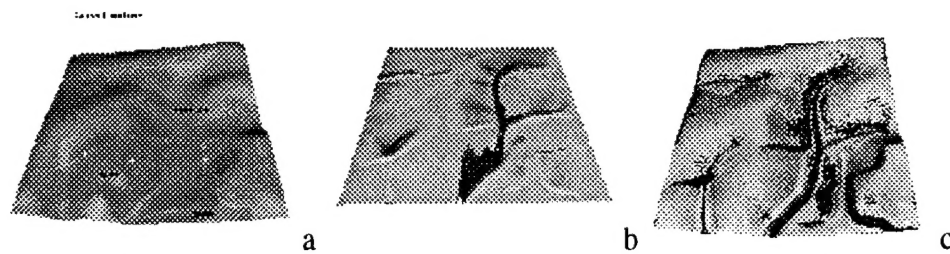


Figure 6. Current situation: a) land use, b) simulated sediment flux, c) erosion/deposition

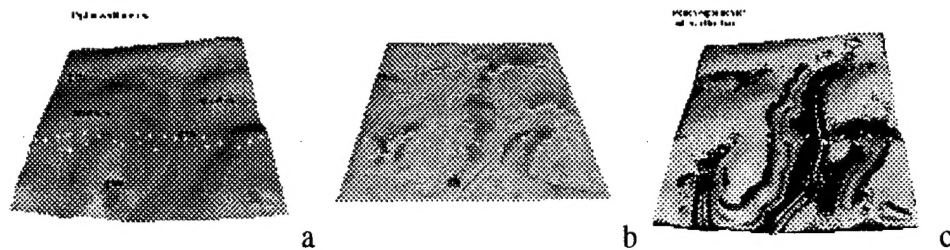


Figure 7. Situation with the grass area spatially distributed for more effective erosion protection : a) land use, b) simulated sediment flux, c) erosion/deposition

CONCLUSION

We have presented an approach to modeling of landscape processes in an advanced GIS environment which is based on the following developments :

- We have used multivariate regularized splines with tension for scattered data interpolations and continuous field representations for a multitude of tasks such as processing of input data, analysis and presentation of simulation results.
- We have developed and employed Monte Carlo methods for solving both water and sediment 2D transport problems. This approach proved to be robust and can be relatively easily generalized to include effects currently omitted (e.g., 3D infiltration process). The stochastic methods are also very well suited for distributed parallel computing.
- Extending the erosion model from traditional 1D water and sediment flow equations to fully 2D fields, supported by the GIS implementation, provided a new insight into the functioning of these models in a complex realistic landscape.
- Fully integrated visualization based on multiple dynamic surfaces helped in various stages of development, evaluation and applications of complex models, where interaction between the spatial fields is important.
- On the basis of these achievements we were able to produce landuse scenario with a potential of significantly decreased erosion in a fully distributed manner.

Important computational components of our approach such as interpolation tool, water flux solver, sediment flux solver, scenario optimizer (under current development) are built as *functional units* with well-defined input, output and controls. Therefore these units can be used either as separate tools or within open GIS frameworks such as GRASS depending on the computational environments and size of tasks.

Using the presented application as an example, we believe that the GIS in future can become not only a

powerful tool for providing and analyzing spatial information, but by extending its capabilities as a simulation and optimization tool, it can allow its users to find unexpected solutions of land management problems leading to practices which can be more effective at lower cost than the currently known conservation approaches.

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